





Seed production and dispersal limit treeline advance in the Pyrenees

Alba Anadon-Rosell¹  | Marc Talavera² | Josep M. Ninot²  | Empar Carrillo²  | Enric Batllori^{2,3} 

¹Institute of Botany and Landscape Ecology, University of Greifswald, Greifswald, Germany

²Department of Evolutionary Biology, Ecology and Environmental Sciences, Biodiversity Research Institute (IRBio), University of Barcelona, Barcelona, Spain

³Centre de Recerca Ecològica i Aplicacions Forestals (CREAF), Universitat Autònoma de Barcelona, Cerdanyola del Vallès, Spain

Correspondence

Alba Anadon-Rosell, Institute of Botany and Landscape Ecology, University of Greifswald, Greifswald, Germany.
Email: alba.anadon-rosell@uni-greifswald.de

Funding information

This work was supported by the project ARBALMONT/786-2012 (Organismo Autónomo Parques Nacionales, Ministerio de Agricultura, Alimentación y Medio Ambiente, Spain) and the Aigüestortes i Estany de Sant Maurici National Park. AAR was supported by a Postdoctoral Research Fellowship from the Alexander von Humboldt Foundation. MT was funded by an FI grant (Generalitat de Catalunya).

Co-ordinating Editor: James M. Bullock

This article is a part of the Special Feature Plant dispersal and establishment as drivers of vegetation dynamics and resilience, edited by Péter Török, James M. Bullock, Borja Jiménez-Alfaro and Judit Sonkoly.

Abstract

Aims: *Pinus uncinata* is the major treeline-forming species in the Pyrenees. Yet, the role of its reproduction and dispersal as drivers of treeline dynamics remains unknown. Here we quantify seed production, dispersal and germination changes along the elevation gradient to assess whether they may constrain the foreseen treeline advance in the Pyrenees.

Location: Central Pyrenees, Catalonia, NE Spain.

Methods: We established four plots along an elevation gradient from the closed sub-alpine forest to the krummholz zone at five study sites. In each plot, we collected cones from five to six trees, measured their length, and triggered their opening in the laboratory to count the number of empty seeds and the number and weight of full seeds. We used the collected seeds in a germination experiment under controlled conditions in growth chambers. Additionally, we installed seed traps along the forest-alpine grassland transition to measure seed rain for three consecutive years in three of the study sites.

Results: The number of full seeds per cone decreased along the elevation gradient and was correlated with cone length. However, the proportion of full seeds per cone and their weight did not differ between elevation positions. Seed rain decreased drastically with elevation and no seeds arrived into the alpine grassland traps consistently across study years. Although germination success did not significantly differ between elevation provenances (i.e., elevation position of origin), we found significant differences in germination dynamics between study sites and between elevation provenances within sites.

Conclusions: Our results indicate that whereas the viability of *Pinus uncinata* seeds is not limited by elevation, seed production and dispersal are constraining the ongoing rates of treeline advance in the Pyrenees.

KEYWORDS

alpine, *Pinus uncinata*, reproductive ecology, seed germination, seed rain, treeline ecotone

This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2019 The Authors. *Journal of Vegetation Science* published by John Wiley & Sons Ltd on behalf of International Association of Vegetation Science

1 | INTRODUCTION

Alpine treelines represent the upper altitudinal limit of the forest in mountain regions, a boundary that is globally constrained by temperature (Körner & Paulsen, 2004). The thermal limitation for forest growth above the treeline explains that the position, dynamics, and structural changes in this ecotone mirror in many cases the effects of climate warming (e.g., Cazzolla Gatti et al., 2019; Harsch, Hulme, McGlone, & Duncan, 2009; Holtmeier & Broll, 2005). However, not only climate but also geomorphological and topographical constraints, as well as anthropogenic activities, determine the current location and physiognomy of treelines across the globe (e.g., Ameztegui, Coll, Brotons, & Ninot, 2016; Holtmeier & Broll, 2005; Leonelli, Pelfini, Cella, & Caravaglia, 2011; Tremblé, Šenfeldr, Chuman, Ponocná, & Demková, 2016). In fact, many reported treeline advancements in elevation may be explained by the decrease and even cessation of anthropogenic activities in the last century. In such cases, treelines are not expanding beyond their historical thermal limits, but recovering their past altitudinal distribution (Ameztegui et al., 2016; Cudlín et al., 2017). Regardless, the treeline advance, either through range expansion or through re-colonisation, may lead to changes in extant plant diversity and ecosystem function through displacement of alpine species and shifts in carbon and nutrient dynamics (Greenwood & Jump, 2014).

Production of viable seeds is the first step for a seed-based regeneration process at any treeline (Holtmeier & Broll, 2005). The next essential step is the dispersion of the seeds, which represents the first spatial demographic process that will determine the position of recruit establishment and potential range shifts (Nathan & Muller-Landau, 2000). Ultimately, any change in treeline position is thus related to successful seed production, viability, dispersal and germination, as well as seedling survival, at higher elevations (or northern latitudes). Some studies have reported a decrease in cone and seed production or seed germination with elevation and latitude (Brown et al., 2019; Juntunen & Neuvonen, 2006; Kruse et al., 2019; Šenfeldr & Maděra, 2011; Sirois, 2000), pointing to potential reproductive limitations at the treeline. These reproductive constraints seem exclusively related to changes in population density in some cases (Juntunen & Neuvonen, 2006), whereas in others they seem related to changes in the reproductive capacity of individual trees (Sirois, 2000), or a combination of both factors (Šenfeldr & Maděra, 2011). However, the reproductive ecology of treelines is still largely unknown (Körner, 2012), especially in relation to the dynamics of alpine treeline ecotones.

In the Pyrenees, decades of study have led to a broad comprehension of ongoing and recent past treeline dynamics (Ameztegui et al., 2016; Angulo, Ninot, Peñuelas, Cornelissen, & Grau, 2019; Batllori, Camarero, & Gutiérrez, 2010; Batllori, Camarero, Ninot, & Gutiérrez, 2009; Batllori & Gutiérrez, 2008; Camarero, Gazol, Galván, Sangüesa-Barreda, & Gutiérrez, 2015; Camarero & Gutiérrez, 2004; Grau, Ninot, Cornelissen, & Callaghan, 2013). The influence of climatic and anthropogenic factors on tree growth, recruitment dynamics, and forest expansion has been thoroughly examined. In

contrast, there is a lack of studies tackling whether and how the reproductive ecology of *Pinus mugo* Turra subsp. *uncinata* (Ramond ex DC.) Domin (hereafter *Pinus uncinata*), the major treeline-forming species in the Pyrenees, may modulate the described patterns and responses to a changing environment (but see Vitali et al., 2019).

In this study, we aimed to address this gap by focusing on seed production, viability, dispersal, and germination dynamics of *Pinus uncinata*. Our main objective was to analyse whether seed production and dispersal can constrain the elevation advancement of the treeline in the Pyrenees, and to determine variations in seed germination dynamics between different provenances (i.e., elevation positions) along the elevation gradient under controlled environmental conditions. We hypothesised that: (a) viable seed production would decrease with elevation; (b) that seed rain and seed arrival would decrease along the elevation gradient; and (c) that seed germination rates and dynamics would mirror the provenance in the elevation gradient, with seeds from lower elevations (where temperatures are higher and growing seasons longer) showing higher germination success and a more extended germination period than seeds from trees growing at higher elevations.

2 | METHODS

2.1 | Study sites and species

Seed production was assessed at five sites in two protected areas of the Central Pyrenees, Catalonia, Spain. The study sites Gelada, Dellui, Son and Eixol are located in the Aigüestortes i Estany de Sant Maurici National Park, and the study site Capifonts is located in the Alt Pirineu Natural Park (Figure 1, Table 1). A sixth study site, Amitges, was used for the seed rain experiment together with Gelada and Dellui (Figure 1, Appendix S1). All the sites occupy NW–NE facing slopes. The anthropogenic alteration of forests in these areas has been minimal since the mid-twentieth century due to their protection status, as well as an ongoing reduction in livestock numbers and the abandonment of forest clearcutting (Domínguez Martín, 2001; Gracia, Meghelli, Comas, & Retana, 2011). Forest expansion has been noticeable from 1956 to the present, although this phenomenon has produced little displacement of the treeline in such near-natural treeline locations (Batllori & Gutiérrez, 2008). The mean annual precipitation in climate stations located near the study sites for the period 2008–2014 is 1,340 mm, with a mean temperature of 9.8°C for the growing season months (JJA) and –3.6°C for the winter months (DJF) (Servei Català de Meteorologia). For detailed climate data, see Appendix S2.

The study species is *Pinus uncinata*, the dominant tree species in most of the treeline ecotones in the Pyrenees (Ninot et al., 2008). It forms dense subalpine forests between 1,700 m and 2,200 m a.s.l. (above sea level), with a potential treeline position that may reach 2,200–2,450 m a.s.l. (Carreras et al., 1996; Ninot et al., 2007). It is a shade-intolerant conifer with a pre-reproductive period of >18 years (Tapias, Climent, Pardos, & Gil, 2004) and masting events every

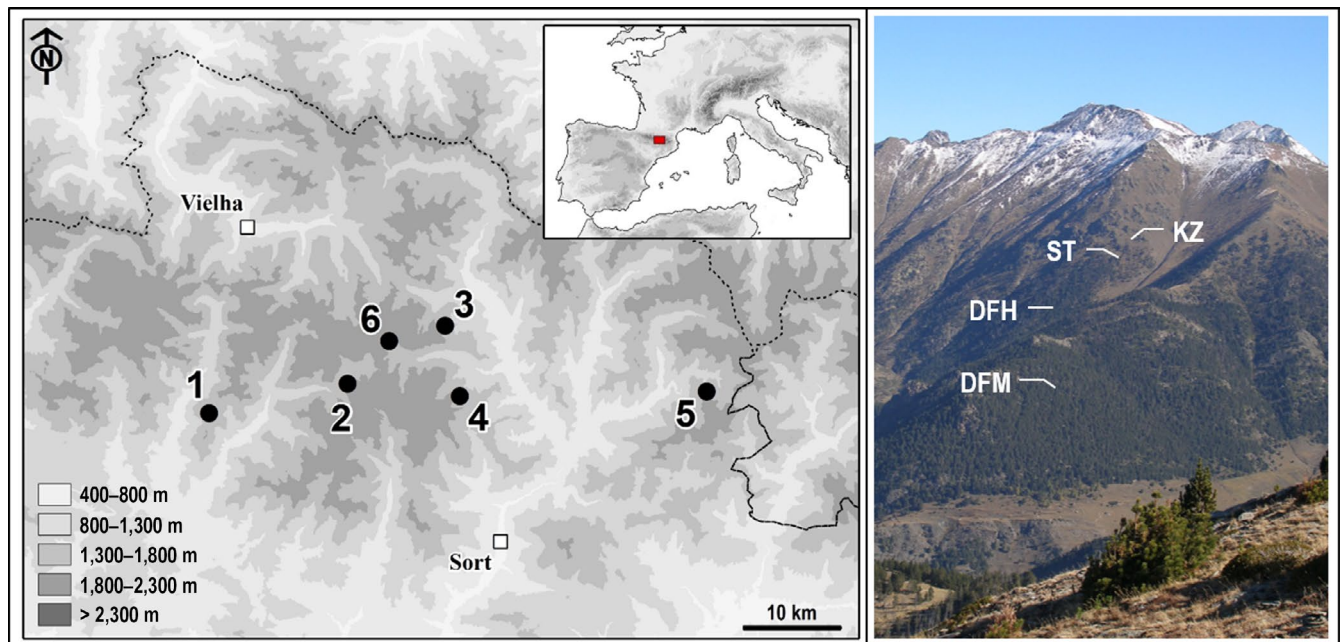


FIGURE 1 (Left) Location of the study sites: 1, Gelada; 2, Dellui; 3, Son; 4, Eixol; 5, Capifonts; 6, Amitges. (Right) Distribution of the sampling plots along the elevation gradient near the study site Capifonts: DFM (dense forest at mid-elevation); DFH (dense forest at high elevation); ST = scattered trees; KZ = krummholz [Colour figure can be viewed at wileyonlinelibrary.com]

3–4 years (Ceballos & Ruiz de la Torre, 1979). Flowering occurs between late spring and the beginning of summer; and seeds reach maturity in late summer of the following year and are dispersed in spring on the third year (Notivol, Oliet, & Serrada, 2012). The winged seeds are wind-dispersed, most frequently over short distances (4–6 m) and less frequently over medium distances (up to 27 m) (Camarero, Gutiérrez, Fortin, & Ribbens, 2005). The favourable germination period runs from late spring (after snowmelt) until the end of summer. It occurs massively at temperatures exceeding 10°C, and it is not significantly affected by variations in the range 15–25°C (Escudero, Pérez-García, & Luzuriaga, 2002).

2.2 | Sampling and data collection

2.2.1 | Seed production along the elevation gradient

We carried out the sampling from mid-October to mid-November 2010. At each of the five study sites, we established a transect along the elevation gradient where we set four plots corresponding to different elevation positions and plant communities: dense forest at mid-elevation within the subalpine belt (DFM), dense forest at maximum elevation within the subalpine belt (DFH), scattered trees (ST), and krummholz zone (KZ) (Figure 1, Table 1). The three highest sampling plots were close to each other and encompass the treeline ecotone. At each plot we sampled 5–6 mature trees and collected 5–13 cones per tree. In total, we collected 635 cones across plots and study sites. The minimum distance between sampled trees was 50 m, and they were selected to include the local forest

heterogeneity (e.g., tree size). To include variability linked to position within the canopy, cones were collected at different orientations across the canopy.

We applied a cold treatment to the collected cones. We placed them in growth chambers for one week at 4°C, one month at –18°C, and 2.5 months at temperatures ranging between –10°C and 9°C, to simulate natural conditions. We triggered cone opening with a moderate heat treatment that consisted of an initial temperature of 30°C that was gradually increased over 8 hr up to 55°C, which was maintained for 10 min. This temperature does not affect the viability of *Pinus uncinata* seeds (Escudero, Sanz, Pita, & Pérez-García, 1999). We mechanically opened those cones that were not completely open after the heat treatment to extract all their seeds. For each cone, we measured the maximum length (from base to apex) and assessed the number of empty seeds (without embryo) and the number and weight of fully developed seeds (hereafter full seeds). Additional information on the assessment of cone production at two of the study sites can be found in Appendix S3.

2.2.2 | Germination experiment

We carried out a germination experiment to test the potential viability and germination dynamics of all full seeds under comparable environmental conditions. We used a growth chamber under controlled conditions at the facilities of Serveis de Camps Experimentals of the University of Barcelona. We set long-day conditions defined by 16 hr light and 8 hr darkness at 20°C and 15°C, respectively. These conditions favour germination of *Pinus uncinata* (Escudero

TABLE 1 Altitude (m a.s.l.) and UTM coordinates (ED50, zone 31T) of each sampling plot at each study site for the seed production and germination analysis

	Gelada (Port de Gelada)		Dellui (Estany de Dellui)		Son (Lo Tésol de Son)		Eixol (Creu de l'Eixol)		Capifonts (Serrat de Capifonts)	
	UTM	Altitude	UTM	Altitude	UTM	Altitude	UTM	Altitude	UTM	Altitude
Dense forest at mid-elevation (DFM)	31,620, 471,221	1,820	33,006, 471,419	1,860	34,099, 471,990	1,925	34,277, 471,280	1,900	36,702, 471,307	2,010
Dense forest at maximum elevation (DFH)	31,637, 471,202	2,000	33,032, 471,394	2,050	34,057, 471,933	2,050	34,266, 471,184	2,100	36,746, 471,227	2,355
Scattered trees (ST)	31,636, 471,178	2,140	33,075, 471,363	2,190	33,996, 471,915	2,345	-	-	36,768, 471,217	2,400
Krummholz (KZ)	31,647, 471,153	2,200	33,110, 471,367	2,300	33,965, 471,917	2,400	34,277, 471,160	2,230	36,768, 471,213	2,430

Note: For each study site, the complete toponym is given in parentheses.

et al., 2002) and are close to those found in the subalpine belt in the most favourable period for germination in early summer (Ninot et al., 2008). We sowed all full seeds from the same cone in one pot (7.5 cm × 7.5 cm × 8 cm), or in two pots when cones had more than 50 full seeds, always at 1-cm depth. We used a sterile substrate of 1:1 perlite and vermiculite, watered weekly to saturation with deionised water and let the pots drain freely. We recorded seedling emergence and removed seedlings twice a week to avoid possible competition or inhibition. The germination trial lasted 121 days.

2.2.3 | Seed rain and seed dispersal along the altitudinal gradient

To explore in more detail seed production and dispersal of *Pinus uncinata*, we performed a seed rain experiment at three of the study sites, Gelada, Dellui and Amitges (Figure 1, Appendix S1). To account for the interannual variability in seed production, we carried out this experiment during three consecutive years. At each site we established one transect with four plots (elevation positions) along the elevation gradient. Plots corresponded to the subalpine dense forest (SDF), subalpine fragmented forest (SFF), krummholz zone (KZ), and alpine grassland (AG). Note that in Gelada and Dellui, seed rain plots do not correspond to the plots used in the seed production experiment (Appendix S1). In Amitges, due to its characteristics, the KZ and AG plots could not be established. Between 28 September and 5 October 2011 we installed six seed traps per plot at each site. Seed traps were 1 m² and were made of artificial tuft grass. In each plot, the six traps were installed forming a circumference of 7–8 m in diameter. From 2012 to 2014 we collected all seeds in each trap during summer and counted them at the laboratory. Since dispersal distances of *Pinus uncinata* are frequently within 10 m from the parental tree (Camarero et al., 2005; Vitali et al., 2019), we established a circular area adding a supplementary radius of 10 m from where the circumference of seed traps was located. Within this circular area (seed trap area + external circumference), we counted all the trees present, we measured their diameter at breast height (DBH), and classified them as reproductive trees (DBH > 5 cm), potentially reproductive trees (non-reproducing sexually mature trees with DBH > 5 cm), non-reproductive trees (DBH < 5 cm), and dead trees. We used these data to calculate reproductive tree density and the mean DBH of reproductive trees in the area covered by the seed traps at each study site and elevation position (see Appendix S4).

2.3 | Statistical analyses

We used linear and generalised linear mixed-effects models to analyse our data. For all the study variables, we performed model selection following the Akaike information criterion (AIC) to select the most suitable model (Zuur, Ieno, Walker, Saveliev, & Smith, 2009). When two models did not differ in more than two AIC units, we applied a model average function (Bartoń, 2019).

We used generalised linear mixed-effects models to assess the effects of the elevation gradient on the number of seeds and number of full seeds per cone (fitted with a Poisson distribution) and on the proportion of full seeds per cone (fitted with a binomial distribution). We used linear mixed-effects models fitted with the restricted maximum-likelihood method (REML) to test the effects of the elevation gradient on cone length and weight of full seeds. For all these variables, we used the elevation position as a fixed-effect factor and site and tree identity as random factors as a starting point for model selection. Additionally, we explored the relationship between cone length and number of full seeds per cone with Pearson correlation tests and exponential regression.

To analyse differences in the germination success (i.e., seed viability) at the end of the germination experiment (seeds germinated out of seeds sown), we used generalised linear mixed-effects models fitted with a binomial distribution. We included elevation position and site as fixed effects and tree identity as a random factor prior to model selection.

For the analysis of germination dynamics, we used meta-analytic random-effects models, which allow incorporating random factors in log-logistic models of germination data (Keshtkar, Mathiassen, Beffa, & Kudsk, 2017; Ritz, Pipper, & Streibig, 2013). For this, we carried a two-step approach. First, we fitted a three-parameter log-logistic model to the cumulative germination data according to Ritz et al. (2013):

$$F(t) = \frac{d}{1 + \exp[b \{ \log(t) - \log(t_{50}) \}]} = \frac{d}{1 + \left(\frac{t}{t_{50}} \right)^b} \quad (1)$$

where $F(t)$ is the cumulative seed germination at time t , d is the upper limit parameter denoting the proportion of seeds that germinated during the experiment out of the total number of seeds sown, t_{50} is the time when 50% of the seeds that germinated during the experiment (d) have germinated, and b is the slope of F at time $t = t_{50}$.

In the second step, we separately fitted the meta-analytic random-effects model to our parameters of interest. Therefore, we analysed the estimates of t_{50} , b and d obtained with the event-time model (first step) with a linear mixed-effect model where tree was defined as a random factor and elevation position, site and the interaction elevation position \times site were defined as fixed effects. Site was included here as a fixed effect to specifically investigate potential effects of the different provenances (i.e., elevation positions) of the seeds on the dynamics of germination. The estimated standard errors for the parameters of interest were also included in the models. We made pairwise comparisons to find significant differences between the factor levels.

We used a zero-inflated negative binomial (ZINB) model to assess the effects of the elevation gradient on seed rain per area. We used these models to account for data overdispersion and zero inflation (their AIC values were lower than in generalised linear models fitted with Poisson and negative binomial distributions). We used site, elevation position, study year and the interactions position \times year and site \times year as fixed-effect factors in the count

part of the model. In the binomial part, we used the three factors without interactions following the model selection procedure. We excluded the alpine grassland plots from all the analyses on seed rain because no seeds were found in these plots at any of the study sites and years.

We tested for normality and homoscedasticity of the residuals and applied data (sqrt- and log-) transformations when necessary to reach these assumptions in the linear regression models. We also used varIdent structure when the homogeneity of residuals was not reached (Zuur et al., 2009). When fixed terms were selected in the final model, we used post-hoc Tukey HSD tests to identify significant differences between fixed term levels.

For all statistical analyses we used R software v. 3.6 (R Core Team, 2018). We used the *LME4* package (Bates, Maechler, Bolker, & Walker, 2015) and the *NLME* package (Pinheiro, Bates, DebRoy, & Sarkar, 2019) for linear mixed-effects models, the *MuMIn* package (Bartoń, 2019) for model average, and the *MULTCOMP* package for post-hoc tests (Hothorn, Bretz, & Westfall, 2008). For the meta-analytic random-effects models on germination data, we used the *DRC* package (event-time model; Ritz, Baty, Streibig, & Gerhard, 2015) and the *METAFOR* package (Viechtbauer, 2010). Finally, we used the package *GGPLOT2* (Wickham, 2016) for graphical display, and the *CREATED-HARMA* function in the *DHARMA* package (Hartig, 2019) to plot the ZINB residuals. A summary of all the final models used for each of the study variables is shown in Table 2.

3 | RESULTS

3.1 | Seed production

The number of total seeds per cone decreased 65% along the elevation gradient from DFM to KZ. Values were significantly lower at KZ (16.4 ± 0.9) than in the other three elevation positions (DFM = 46.8 ± 1.8 , DFH = 36.6 ± 1.7 , ST = 33.3 ± 1.7 , Tukey post-hoc test $p < 0.05$). The number of full seeds per cone also decreased along the gradient, from 37.5 ± 1.8 full seeds per cone at DFM to 14.0 ± 0.8 at KZ (Table 3, Figure 2a), which represents a 63% reduction. Post-hoc tests showed significant differences between KZ and the forest sites (DFM and DFH, $p < 0.05$) but not between KZ and ST ($p = 0.171$). The elevation position showed no significant effect on the proportion of full seeds to empty seeds per cone and on the weight of the full seeds (mean across sites and elevation positions of 9.15 ± 0.2 mg/seed) (Table 3).

The elevation gradient also had an effect on cone length, which decreased with increasing elevation (Table 3). Post-hoc tests showed that at the two higher elevations, KZ (3.92 ± 0.05 cm) and ST (4.35 ± 0.06 cm), cone length was significantly lower than in the two forest plots, and within these two, the lower subalpine forest DFM (4.79 ± 0.05 cm) showed longer cones than the higher subalpine forest DFH (4.43 ± 0.06 cm). Correlation analyses showed a significant exponential correlation between cone length and total full seeds per cone (no. viable seeds = exp

**TABLE 2** Explanatory models selected following the AIC criteria for each of the study variables

Variable	Fixed factors	Random factors	Model and error distribution		Explanatory power
No. seeds per cone	Gradient	Gradient Site/Tree	glmm	Poisson	R^2_m : 0.34, R^2_c : 0.97
No. full seeds per cone	Gradient	Gradient Site/Tree	glmm	Poisson	R^2_m : 0.27, R^2_c : 0.96
Proportion of full seeds per cone	–	Site/Tree	glmm	Binomial	R^2_m : –, R^2_c : 0.19
Weight of full seeds	–	Tree	lme	Gaussian	R^2_m : –, R^2_c : 0.53
Cone length	Gradient	Site/Tree	lme	Gaussian	R^2_m : 0.14, R^2_c : 0.62
Total germination (model average)	Site	Tree	glmm	Binomial	R^2_m : 0.02, R^2_c : 0.14
	Site × Gradient	Tree	glmm	Binomial	R^2_m : 0.05, R^2_c : 0.14
Germination dynamics (t_{50})	Gradient × Site	Tree	Meta-Analytic Random Effects		Pseudo- R^2 : 0.30
Seed rain per area	(Site + Gradient) × Year	–	zinb	Negative binomial (zero-inflated)	Pseudo- R^2 (Nagelkerke) 0.75

Note: When the function model average was applied, the two models are shown.

R^2_m and R^2_c for lme and glmm models calculated with the *r.squaredGLMM* function (MuMIn package in R) (for binomial models, the theoretical values are shown). Pseudo- R^2 for the meta-analytic random-effects model calculated as the proportional reduction in the variance components [(variance of the model fitted without coefficients – variance of the model fitted with coefficients)/ variance of the model fitted without coefficients]. Pseudo- R^2 for zinb model calculated with the function *nagelkerke* (rcompanion package): [(null deviance – residual deviance)/null deviance].

($0.67 + 0.57 \times \text{cone length}$), $p < 0.001$ for both parameters; Figure 2b), and also with the total number of seeds per cone (no. seeds = $\exp(1.06 + 0.54 \times \text{cone length})$), $p < 0.001$ for both parameters). Within a given position along the elevation gradient, however, the exponential relationship between cone length and seed production was only maintained at DFH, whereas it was linear at the other elevation positions (Appendix S5).

3.2 | Seed germination

A mean of 65% of the sown seeds across the four assessed elevation positions germinated after 121 days (Figure 3, Appendix S6). For the analysis of the germination success (proportion of germinated seeds at the end of the experiment) we selected two generalised linear mixed-effects models with a difference in AIC lower than two units. Elevation, study site, and their interaction were included in one model whereas only site was included in the other. Model average only showed a significant effect of study site on the germination success (Table 4, Appendix S6), with the largest differences between Dellui and Eixol (post-hoc Tukey test run for each model, $p < 0.01$).

The meta-analytic random-effects models did not show consistent patterns in the germination dynamics between elevation plots across study sites for any of the estimated parameters (Figure 3). However, when looking at the study sites independently we found significant differences between elevation positions (Appendix S7). We focus here on t_{50} to illustrate the time (number of days) when 50% of the seeds at a given elevation had germinated; for results on parameters b and d see Appendix S8. The largest differences in t_{50} between elevation

plots occurred in Gelada, where the forest at high altitude (DFH) had a lower t_{50} than the lower forests (DFM; estimated difference of 29.3 ± 11.7 days, $p = 0.012$) and in Son, where the krummholz zone (KZ) had a lower t_{50} than DFH (estimated difference of 26.6 ± 11.9 days, $p = 0.025$). At the study site Capifonts, the scattered trees (ST) had a higher t_{50} than (KZ; estimated difference of 25.3 ± 9.7 days, $p = 0.009$) and higher than DFM (estimated difference 25.2 ± 11.6 days, $p = 0.030$). Finally, Dellui KZ had a lower t_{50} than DFM (estimated difference 21.9 ± 11.2 days, $p = 0.05$, respectively).

3.3 | Seed rain and dispersal

Seed rain significantly decreased along the elevation gradient. The number of seeds recovered at each elevation position was significantly lower than at the elevation directly below within the elevation gradient, and it was zero at alpine grasslands (Table 5, Figure 4, $p < 0.001$ after Tukey post-hoc tests between each elevation plot). The study site and the study year also had a significant influence on seed rain values. The number of seeds recovered in the seed traps was significantly larger at Amitges (44.30 ± 8.59) than at Gelada (15.06 ± 4.05 , $p < 0.001$ after Tukey post-hoc test) and at Gelada than at Dellui (22.15 ± 4.30 , $p = 0.035$ after Tukey post-hoc test) across all years and elevation plots. Seed rain was significantly larger in 2012 (50.79 ± 8.87) than in the other two years (in 2013 15.47 ± 2.81 , and in 2014 15.38 ± 3.88 , $p < 0.001$ after Tukey post-hoc test in both cases) across sites and elevation positions. In addition, we found significant interactions between elevation and year and between site and year (Table 5), but the patterns showed the same direction in all cases (Appendix S9).

**TABLE 3** Results of the statistical models for the parameters related to seed production

Parameter	Fixed effects	Estimate	SE	z/t-value	p	Random effects		Variance	St. Dev.
No. seeds/cone	Intercept	3.774	0.161	23.441	<0.001	Tree:Site	Intercept	0.080	0.284
	GradientDFH	-0.266	0.126	-2.103	0.035		GradientDFH	0.115	0.339
	GradientST	-0.428	0.364	-1.175	0.240		GradientST	0.038	0.195
	GradientKZ	-1.159	0.212	-5.474	<0.001		GradientKZ	0.347	0.589
						Site	Intercept	0.114	0.338
							GradientDFH	0.005	0.072
							GradientST	0.59	0.768
							GradientKZ	0.138	0.372
No. full seeds/cone	Intercept	3.518	0.201	17.48	<0.001	Tree:Site	Intercept	0.105	0.324
	GradientDFH	-0.269	0.155	-1.735	0.083		GradientDFH	0.320	0.566
	GradientST	-0.453	0.413	-1.096	0.273		GradientST	0.049	0.222
	GradientKZ	-1.083	0.251	-4.311	<0.001		GradientKZ	0.545	0.738
						Site	Intercept	0.183	0.427
							GradientDFH	0.033	0.183
							GradientST	0.771	0.878
							GradientKZ	0.219	0.468
Proportion of full seeds/cone	Intercept	1.652	0.157	10.530	<0.001	Tree:Site	Intercept	0.692	0.832
						Site	Intercept	0.085	0.292
Weight full seeds	Intercept	9.149	0.190	48.120	<0.001	Tree	Intercept	3.206	1.791
Cone length	Intercept	4.810	0.130	37.095	<0.001	Tree:Site	Intercept	0.299	0.547
	GradientDFH	-0.382	0.157	-2.429	0.017	Site	Intercept	0.022	0.150
	GradientST	-0.433	0.173	-2.502	0.014				
	GradientKZ	-0.838	0.158	-5.304	<0.001				

Note: The estimated coefficient and standard error, z- or t-value, and p-value are shown for the fixed terms, including the interaction between two fixed terms when selected in the final model. The estimated variance and standard deviation are shown for the random effects in the mixed-effects models.

4 | DISCUSSION

Our study investigated the reproductive and dispersal patterns of the treeline-forming species *Pinus uncinata* along the treeline ecotone in the Pyrenees. Our findings on seed production, viability, and dispersal provide evidence of reproduction as a limiting factor for treeline advance.

4.1 | Seed production and viability at the treeline

Our results show a clear decrease in seed production of *Pinus uncinata* along the elevation gradient from the dense subalpine forest to the krummholz zone. Our analysis suggests that such decrease is directly related to the size of the cones.

Heat sum fluctuations, thermal limitations, and shorter growing seasons at high elevations may affect both cone production and development, and may have negative effects on fertilisation and embryo growth processes (Almqvist, Bergsten, Bondesson, & Eriksson, 1998; Sirois, 2000). Climate may act in synergy with other factors such as forest structure, resulting in lower pollen availability and an increased self-fertilisation (selfing) with the decrease in

forest density along the elevation gradient (Iwasaki et al., 2013; Smith, Hamrick, & Kramer, 1988). We did not find a reduction in the proportion of full seeds with elevation, which would be the first negative outcome of selfing in other *Pinus* species (Iwasaki et al., 2013). Contrarily, our results suggest that seed development on the *Pinus uncinata* alpine treeline could be more influenced by low pollen supply at higher elevations (Iwaizumi & Takahashi, 2012). The decrease in the production of full seeds found along the gradient in our study sites is not accompanied by a reduction in the number of cones produced per tree (see Appendix S3), nor in the weight of the seeds or in their viability. This suggests that the effect of harsher conditions at the upper treeline on reproductive output may constrain pollen production, pollen viability and pollination, as well as investments in cone growth.

Previous studies have reported the influences of climate and forest structure on the reproduction of other treeline species. Sirois (2000) found a significant decrease in the number of seeds and the percentage of filled seeds of black spruce (*Picea mariana*) in a boreal forest treeline transect in northern Québec, Canada. This author did not find differences in cone production along the transect (similar to our results), and suggested that the decrease in seed production and viability was related to a low viability of the pollen at the treeline.

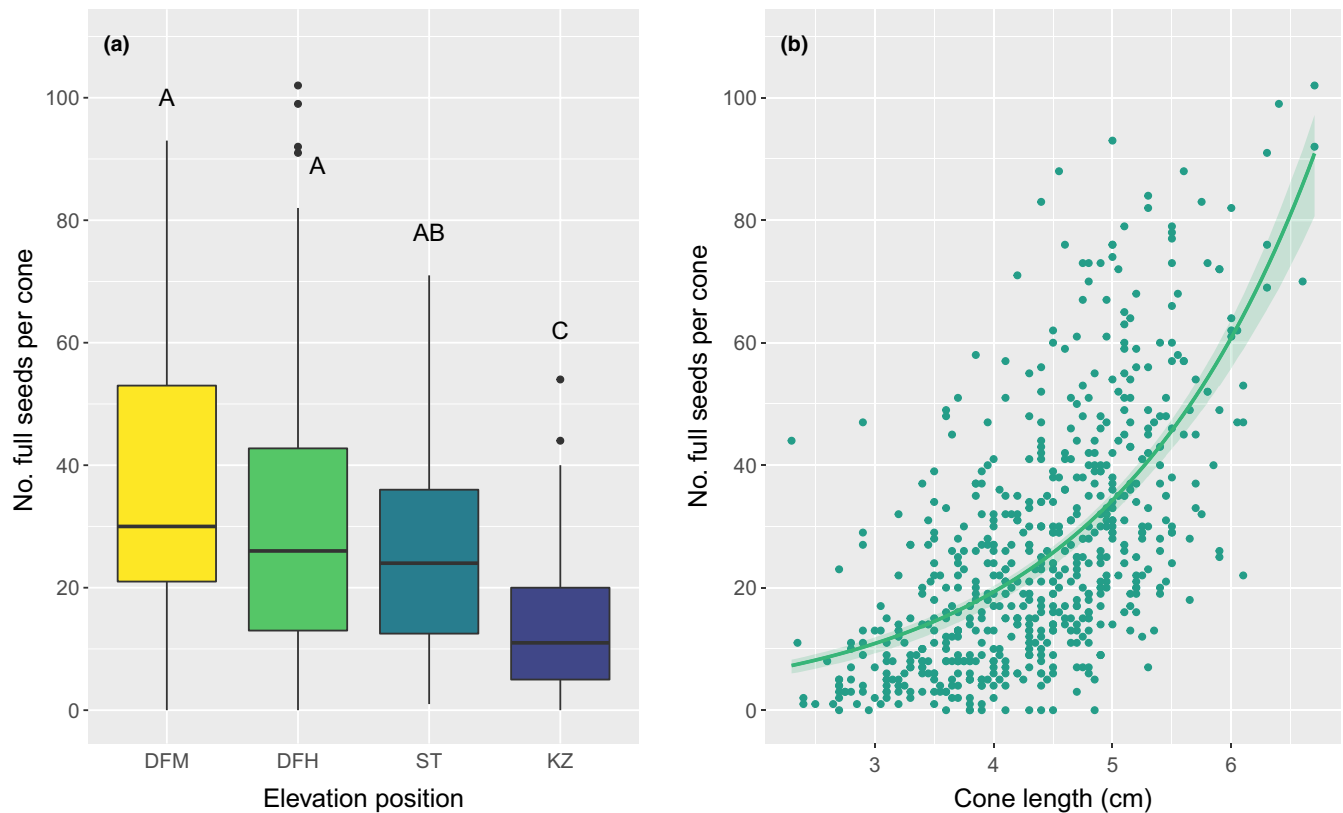


FIGURE 2 (a) Production of full seeds in *Pinus uncinata* cones along the elevation gradient. Capital letters indicate significant differences between elevation positions. DFM = dense forest at mid-elevation; DFH = dense forest at maximum elevation; ST = scattered trees; KZ = krummholz. (b) Relationship between number of full seeds per cone and cone length. Dots correspond to individual values, the green line depicts an exponential regression fit, and the shaded area indicates the 95% confidence interval (residual standard error = 16.51). See correlation plots between cone length and number of full seeds for each elevation position separately in Appendix S5 [Colour figure can be viewed at wileyonlinelibrary.com]

Šenfeldr and Maděra (2011) also found a clear decrease in the reproductive output of Norway spruce (*Picea abies*) in a former pastoral timberline ecotone in the Czech Republic, namely fewer seeds per cone, fewer cones per ha, and fewer cones per fertile specimen at the upper part of the ecotone than at the lower part. Another study in northern Finland found slight differences in seed anatomical maturity between timberline and treeline populations for Scots pine (*Pinus sylvestris*) but no differences in the number of cones per tree (Juntunen & Neuvonen, 2006). Overall, it seems a common pattern to find a decrease in reproductive output related to seed set of the treeline-forming species across elevation or latitudinal gradients, whereas patterns related to cone production and the proportion of viable seeds produced seem more variable among species.

Our findings strongly suggest that the production of viable seeds is the first limiting factor for forest expansion at the Pyrenean treeline. Even under the ongoing climate warming (IPCC, 2018), and unlike the situation in other treeline-forming species (Kullman, 2002), this suggests that the reproduction of *Pinus uncinata* at the treeline is still strongly limited by the present climate conditions. We acknowledge that fluctuations in the reproductive output due to inter-annual climate variations may occur (see section 4.2, *Seed rain and dispersal*, below), but we believe that the reported patterns and the relationship between the parameters studied (e.g., proportion

of full seeds) would not be altered by including additional sampling years. The projected temperature increase could enhance reproduction at high elevations by the alleviation of low temperatures in such temperature-limited ecosystems (Körner, 2012), but forecasted drier conditions in the Pyrenees (López-Moreno & Beniston, 2008) could hamper the effects of increased temperatures.

Contrary to our expectations, we did not find a significant effect of elevation provenance on seed viability (i.e., germination success) and germination dynamics across sites, although the influence of elevation was marginally observed when assessing the sites separately. The lack of a consistent effect on the germination success by the elevation provenance reported here contrasts with other studies showing that germination is an adaptive trait that can vary largely with elevation (Castanha, Torn, Germino, Weibel, & Kueppers, 2013; Rehfeldt, 1989). Sirois (2000) showed that, 30 days after sowing, the percentage of total germination from seeds originating in trees closer to the treeline was lower than in those from the dense boreal forest. Results of Šenfeldr and Maděra (2011) reflected a similar pattern after 21 days, but they did not find significant differences. The different study species and the longer length of our germination trial (121 days) make it difficult to compare between these studies and our experiment. We did find, however, large differences in the germination success of seeds from

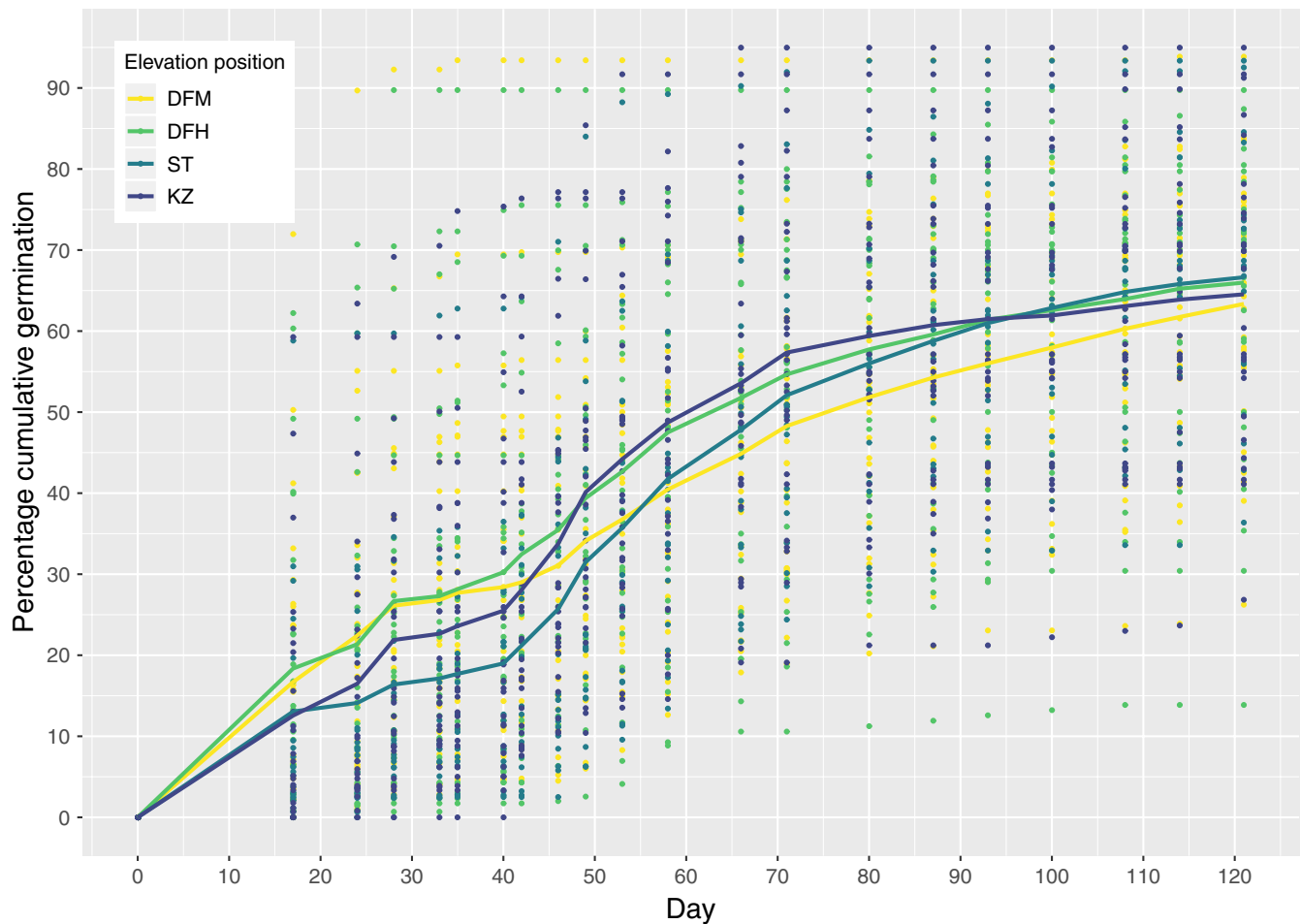


FIGURE 3 Percentage of germinated seeds (germinated seeds/sown seeds \times 100) through time. Lines show the mean across trees and sites for each elevation position. Dots correspond to each individual tree (average of all cones per tree). For more detailed data on germination dynamics at each study site and elevation position see Appendix S7. DFM = dense forest at mid-elevation; DFH = dense forest at high elevation, ST = scattered trees, KZ = krummholz [Colour figure can be viewed at wileyonlinelibrary.com]

different study sites. This suggests that there are stronger genetic barriers between sites than within sites, which may be explained by the gene flow along the elevation gradient being less constrained by altitudinal winds than by topographic barriers between neighbouring valleys. Indeed, topography can be a considerable barrier for gene flow (Ohsawa & Ide, 2008) and populations from neighbouring valleys may show local adaptations that differentiate them from each other (Di Pierro et al., 2017). It is thus important to consider several study areas when trying to find general patterns in seed-based regeneration at the treeline. Further analyses should determine the extent of local adaptation within sites vs. the extent of seed germination plasticity in *Pinus uncinata*.

4.2 | Seed rain and dispersal

The seed rain experiment showed large differences between study sites and study years, emphasising the importance of local and inter-annual conditions for the reproduction dynamics of *Pinus uncinata*. However, despite the seed rain being much larger

in 2012 than in the following years (2013 and 2014), we found a clear decrease in the number of seeds recovered in the seed traps along the elevation gradient that was constant through time. No seeds were found in the alpine grassland beyond the treeline, indicating that seed arrival, which is the first step for an eventual forest colonisation of alpine grasslands, is a limiting factor for treeline advance. Although the area covered by our seed traps is relatively small for concluding that no seeds arrive at alpine grasslands, our findings strongly suggest that important limitations for seed arrival beyond the krummholz zone exist. We acknowledge the possibility that seed predation influences our results, since it has been found to largely affect regeneration beyond the upper elevation limit in some species (Brown & Vellend, 2014; Castro, 1999). However, results from an extra plot for seed traps in a grassland area below the krummholz zone at Amitges showed the recovery of seeds in the traps (0–4 seeds in most of the traps in 2012–2014, and up to 20 seeds in one trap in 2012). This validates the seed trap methodology in grassland areas, and lines up with our reasoning that we did not capture seeds in the alpine grasslands because their arrival there is much more limited as a result



Fixed effects	Estimate	SE	Adjusted SE	z-value	Pr(> z)
(Intercept)	1.194	0.237	0.237	5.034	<0.001
GradientDFH	-0.164	0.307	0.307	0.533	0.594
GradientDFM	-0.319	0.393	0.394	0.811	0.417
GradientKZ	0.312	0.425	0.425	0.733	0.464
SiteCapifonts	-0.746	0.348	0.348	2.143	0.032
SiteEixol	-1.247	0.484	0.484	2.574	0.010
SiteGelada	-0.078	0.343	0.343	0.228	0.820
SiteSon	-0.710	0.356	0.356	1.993	0.046
GradientDFH:SiteCapifonts	0.400	0.515	0.516	0.775	0.439
GradientDFM:SiteCapifonts	0.089	0.407	0.408	0.217	0.828
GradientKZ:SiteCapifonts	0.080	0.444	0.445	0.181	0.857
GradientDFH:SiteEixol	0.232	0.480	0.481	0.483	0.629
GradientDFM:SiteEixol	0.787	0.782	0.783	1.006	0.315
GradientDFH:SiteGelada	0.117	0.408	0.409	0.287	0.774
GradientDFM:SiteGelada	-0.051	0.408	0.409	0.125	0.901
GradientKZ:SiteGelada	-0.734	0.749	0.749	0.980	0.327
GradientDFH:SiteSon	0.160	0.427	0.427	0.374	0.708
GradientDFM:SiteSon	0.756	0.748	0.749	1.009	0.313
GradientKZ:SiteSon	-0.493	0.614	0.615	0.803	0.422

TABLE 4 Results of the model average (full average) for germination success

Note: The estimated coefficient and standard error, z-value, and p-value are shown for the fixed terms and the interaction between the two fixed terms. Full average means that model average is calculated by assuming that all the variables are included in both models, but the corresponding coefficient (and variance) is set to zero in the model where the variable was originally absent (Bartoń, 2019).

of dispersal constraints for *Pinus uncinata* seed (Camarero et al., 2005; Vitali et al., 2019).

The decrease in seed recovery with elevation was not significantly correlated with reproductive tree density, although the number of reproductive trees per area also decreased along the elevation gradient (Appendix S4). We believe these results need to be taken cautiously because there was a large variation in tree density among sites in the two subalpine forest positions (SFF and SDF), and the forest structure of the study sites was complex. There is, however, a clearly lower density of reproductive trees in the krummholz zone, which suggests that the lower number of seeds per area recovered there may be driven both by a lower reproductive capacity of individual trees and a decrease in tree density, as reported in other studies (Juntunen & Neuvonen, 2006; Šenfeldr & Maděra, 2011). Contrarily, the decrease in seed recovery from the subalpine dense forest (SDD) to the subalpine fragmented forest (SFF) may be mostly driven by a decrease in the reproductive capacity of individual trees, as shown by Sirois (2000). We found a positive correlation between the mean DBH of reproductive trees and the number of seeds recovered per area (Appendix S4), similarly to results found by Vitali, Camarero, Garbarino, Piermattei, and Urbinati (2017) on cone production. This reinforces the notion that the performance and characteristics of individual trees play an important role in explaining the reproductive output at treeline regardless of the forest structure.

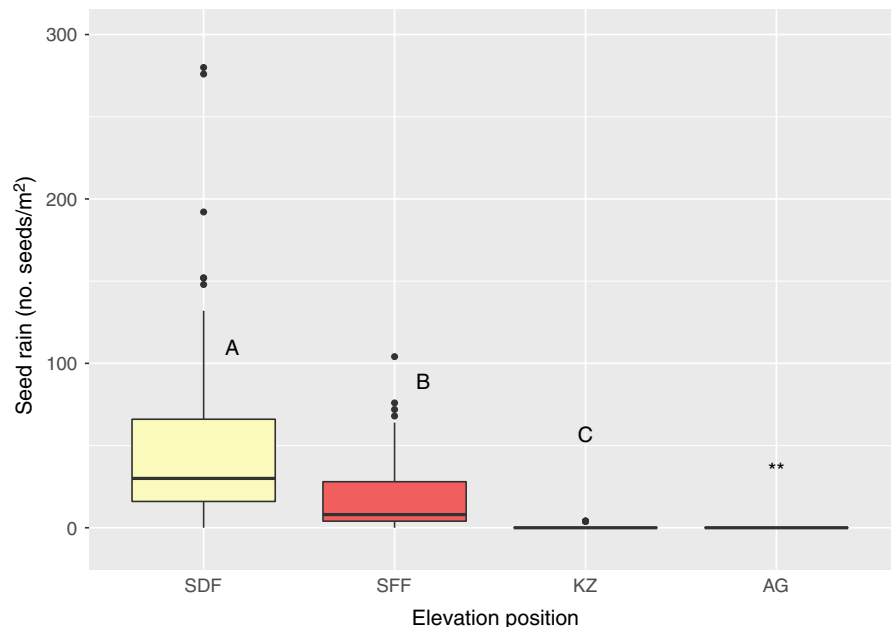
We suggest that seed dispersal is of utmost importance in limiting treeline advance in the Pyrenees. Forest movement upslope will mainly depend on the establishment of seedlings germinating from nearby, scattered trees located at high elevation and rarely from long-distance dispersed seeds (Camarero et al., 2005; Kruse et al., 2019; Nathan, 2006). This may explain why, at the regional scale, ecotone densification (i.e., between the forest and the tree limit) is a more common response to global change in the Pyrenean treelines (Batllori et al., 2010; Batllori & Gutiérrez, 2008) than treeline advance. Johnson, Gaddis, Cairns, and Krutovsky (2017) argued that seed dispersal was extensive enough in mountain hemlock (*Tsuga mertensiana*) in Alaska to ensure the advancement to higher elevations. However, this species has an average dispersal distance of 73 m, and has a long-distance dispersal capacity of up to 450 m, values much greater than those reported for *Pinus uncinata* (Camarero et al., 2005; Vitali et al., 2019). Therefore, the biology of the treeline-forming species may be of paramount importance in their reproductive dynamics, precluding any generalisation in terms of the reproduction-related limiting factors for treeline advance at the global scale. Additionally, seedling recruitment and the eventual treeline advance will not only depend on the arrival of the seeds, but also on the subsequent biotic and abiotic processes affecting seedling establishment and survival (Lett & Dorrepaal, 2018; Nathan & Muller-Landau, 2000), such as the environmental harshness and scarcity of favourable sites for establishment (Smith, Germino,



TABLE 5 Results of the zero-inflated negative binomial model for seed rain

	Estimate	SE	z-value	Pr(> z)
Count model coefficients				
Intercept	5.055	0.213	23.768	<0.001
SiteDellui	-0.328	0.256	-1.280	0.201
SiteGelada	-1.967	0.315	-6.236	<0.001
GradientSFF	-1.065	0.238	-4.485	<0.001
GradientKZ	-3.413	0.452	-7.550	<0.001
Year2013	-1.303	0.316	-4.126	<0.001
Year2014	-2.329	0.366	-6.363	<0.001
SiteDellui:Year2013	0.024	0.387	0.061	0.951
SiteGelada:Year2013	1.537	0.463	3.320	<0.001
SiteDellui:Year2014	0.382	0.393	0.972	0.331
SiteGelada:Year2014	3.093	0.484	6.387	<0.001
GradientBSF:Year2013	-0.165	0.352	-0.467	0.640
GradientLA:Year2013	1.109	120.228	0.923	0.356
GradientBSF:Year2014	0.860	0.379	2.267	0.023
GradientLA:Year2014	-14.975	1,145.702	-0.013	0.990
Log(theta)	0.792	0.155	5.108	<0.001
Zero-inflation model coefficients				
Intercept	-23.768	3,665.699	-0.006	0.995
SiteDellui	1.327	1.428	0.929	0.353
SiteGelada	3.698	1.378	2.684	0.007
Year2013	1.444	0.813	1.776	0.076
Year2014	0.619	1.037	0.597	0.550
GradientSFF	19.696	3,665.698	0.005	0.996
GradientKZ	21.700	3,665.699	0.006	0.995

FIGURE 4 Seeds collected in the seed traps along the elevation gradient (SDF = subalpine dense forest; SFF = subalpine fragmented forest; KZ = krummholz; AG = alpine grassland). Data are plotted across the three study sites (Dellui, Amitges and Gelada) and the three study years (2012–2014). Capital letters indicate significant differences between elevation positions. The asterisk for AG indicates that no seeds were found in any of the seed traps, and thus it was excluded from the analysis. See Appendix S9 for the number of seeds recovered in the seed traps at each elevation position per study year and site [Colour figure can be viewed at wileyonlinelibrary.com]



Hancock, & Johnson, 2003), seed and seedling predation (Cairns & Moen, 2004; Speed, Austrheim, Hester, & Myrsterud, 2010), and facilitation and competition with coexisting mosses, shrubs and herbaceous species (Dullinger, Dirnböck, & Grabherr, 2004; Lett, Wardle, Nilsson, Teuber, & Dorrepaal, 2018; Liang et al., 2016). Former

studies on *Pinus uncinata* at nearby study areas showed high seedling mortality rates above the tree limit when seedlings were planted without the protection of krummholz or shrubs (Batllori et al., 2009), and clear positive effects of shrubs on seedlings through their protection against winter damage and browsing (Grau et al., 2013).



5 | CONCLUSIONS

Our study provides the first evidence of reproductive output as a limiting factor for the advance of the treeline in the Pyrenees. The decrease in seed production along the elevation gradient together with a lack of seed arrival at higher elevations (i.e., alpine grassland) strongly suggests that seed production and dispersal are constraining the ongoing rates of treeline advance in the Pyrenees. By contrast, the lack of clear differences in germination success and germination dynamics from the subalpine forests to the krummholz zone seems to indicate a limited role of genetic constraints at the site level, although genetic variability may have a more important role in treeline dynamics at the regional scale (differential site dynamics). Overall, despite potential positive effects of warming on seed production, our findings suggest that the poor dispersal capacity of *Pinus uncinata* seeds to the alpine grasslands beyond the treeline will slow down the upslope advancement of the forest. This, together with other climate change-related events such as increased droughts, which may imply higher seedling mortality rates, may reduce the expected climatic sensitivity of treeline position to warming climates.

ACKNOWLEDGEMENTS

We would like to thank Christian Ritz, from the University of Copenhagen, for his statistical advice on meta-analytic random-effects models. We are grateful to Albert Ferré for providing the map and the picture in Figure 1. We also thank Oriol Grau, Francesc Talavera, Cecília Roma, Xavier Jou, Gerard Fuentes, Andreu Passola, Enric Cuadras and Xavier Trullàs for their help in the field. Finally, we are grateful to Matteo Garbarino and an anonymous reviewer for their helpful comments on the manuscript.

AUTHOR CONTRIBUTIONS

MT, JMN, EC and EB conceived the research idea. MT and JMN collected cones, seeds and data and performed the laboratory work. AAR performed the statistical analyses. AAR wrote the manuscript with contributions from MT and EB. All the authors discussed the results and commented on the manuscript.

DATA AVAILABILITY STATEMENT

Data supporting this publication is stored at https://github.com/anadon-rosell/Anadon-Rosell_et_al_2019_JVS.

ORCID

Alba Anadon-Rosell  <https://orcid.org/0000-0002-9447-7795>

Josep M. Ninot  <https://orcid.org/0000-0002-3712-0810>

Empar Carrillo  <https://orcid.org/0000-0002-4016-3520>

Enric Batllori  <https://orcid.org/0000-0002-2130-0489>

REFERENCES

- Almqvist, C., Bergsten, U., Bondesson, L., & Eriksson, U. (1998). Predicting germination capacity of *Pinus sylvestris* and *Picea abies* seeds using temperature data from weather stations. *Canadian Journal of Forest Research*, 28(10), 1530–1535. <https://doi.org/10.1139/x98-139>
- Ameztegui, A., Coll, L., Brotons, L., & Ninot, J. M. (2016). Land-use legacies rather than climate change are driving the recent upward shift of the mountain tree line in the Pyrenees. *Global Ecology and Biogeography*, 25(3), 263–273. <https://doi.org/10.1111/geb.12407>
- Angulo, M. A., Ninot, J. M., Peñuelas, J., Cornelissen, J. H. C., & Grau, O. (2019). Tree sapling responses to 10 years of experimental manipulation of temperature, nutrient availability, and shrub cover at the Pyrenean treeline. *Frontiers in Plant Science*, 9, 1871. <https://doi.org/10.3389/fpls.2018.01871>
- Bartoń, K. (2019). *MuMIn: Multi-model inference. R package version 1.43.6*. Retrieved from <https://CRAN.R-project.org/package=MumIn>
- Bates, D., Maechler, M., Bolker, B., & Walker, S. (2015). Fitting linear mixed-effects models using lme4. *Journal of Statistical Software*, 67(1), 1–48. <https://doi.org/10.18637/jss.v067.i01>
- Batllori, E., Camarero, J. J., & Gutiérrez, E. (2010). Current regeneration patterns at the tree line in the Pyrenees indicate similar recruitment processes irrespective of the past disturbance regime. *Journal of Biogeography*, 37(10), 1938–1950. <https://doi.org/10.1111/j.1365-2699.2010.02348.x>
- Batllori, E., Camarero, J. J., Ninot, J. M., & Gutiérrez, E. (2009). Seedling recruitment, survival and facilitation in alpine *Pinus uncinata* tree line ecotones. Implications and potential responses to climate warming. *Global Ecology and Biogeography*, 18(4), 460–472. <https://doi.org/10.1111/j.1466-8238.2009.00464.x>
- Batllori, E., & Gutiérrez, E. (2008). Regional tree line dynamics in response to global change in the Pyrenees. *Journal of Ecology*, 96(6), 1275–1288. <https://doi.org/10.1111/j.1365-2745.2008.01429.x>
- Brown, C. D., Dufour-Tremblay, G., Jameson, R. G., Mamet, S. D., Trant, A. J., Walker, X. J., ... Johnstone, J. F. (2019). Reproduction as a bottleneck to treeline advance across the circumpolar forest tundra ecotone. *Ecography*, 42(1), 137–147. <https://doi.org/10.1111/ecog.03733>
- Brown, C. D., & Vellend, M. (2014). Non-climatic constraints on upper elevational plant range expansion under climate change. *Proceedings of the Royal Society B: Biological Sciences*, 281, 20141779. <https://doi.org/10.1098/rspb.2014.1779>
- Cairns, D. M., & Moen, J. (2004). Herbivory influences tree lines. *Journal of Ecology*, 92(6), 1019–1024. <https://doi.org/10.1111/j.1365-2745.2004.00945.x>
- Camarero, J. J., Gazol, A., Galván, J. D., Sangüesa-Barreda, G., & Gutiérrez, E. (2015). Disparate effects of global-change drivers on mountain conifer forests: Warming-induced growth enhancement in young trees vs. CO₂ fertilization in old trees from wet sites. *Global Change Biology*, 21(2), 738–749. <https://doi.org/10.1111/gcb.12787>
- Camarero, J. J., & Gutiérrez, E. (2004). Pace and pattern of recent treeline dynamics: Response of ecotones to climatic variability in the Spanish Pyrenees. *Climatic Change*, 63(1), 181–200. <https://doi.org/10.1023/B:CLIM.0000018507.71343.46>
- Camarero, J. J., Gutiérrez, E., Fortin, M.-J., & Ribbens, E. (2005). Spatial patterns of tree recruitment in a relict population of *Pinus uncinata*: Forest expansion through stratified diffusion. *Journal of Biogeography*, 32(11), 1979–1992. <https://doi.org/10.1111/j.1365-2699.2005.01333.x>
- Carreras, J., Carrillo, E., Masalles, R. M., Ninot, J. M., Soriano, I., & Vigo, J. (1996). Delimitation of the supra-forest zone in the Catalan Pyrenees. *Bulletin de la Société Linnéenne de Provence*, 47, 27–36.
- Castanha, C., Torn, M. S., Germino, M. J., Weibel, B., & Kueppers, L. M. (2013). Conifer seedling recruitment across a gradient from forest to alpine tundra: Effects of species, provenance, and site. *Plant Ecology and Diversity*, 6(3–4), 307–318. <https://doi.org/10.1080/17550874.2012.716087>
- Castro, J. (1999). Seed mass versus seedling performance in Scots pine: A maternally dependent trait. *New Phytologist*, 144(1), 153–161. <https://doi.org/10.1046/j.1469-8137.1999.00495.x>



- Cazzolla Gatti, R., Callaghan, T., Velichevskaya, A., Dudko, A., Fabbio, L., Battipaglia, G., & Liang, J. (2019). Accelerating upward treeline shift in the Altai Mountains under last-century climate change. *Scientific Reports*, 9(1), 7678. <https://doi.org/10.1038/s41598-019-44188-1>
- Ceballos, L., & Ruiz de la Torre, J. (1979). *Árboles y arbustos de la España Peninsular*. Madrid, Spain: E.T.S.I.M.
- Cudlín, P., Cudlín, P., Cudlín, P., Tognetti, R., Malis, F., Alados, C. L., ... Wielgolaski, F. E. (2017). Drivers of treeline shift in different European mountains. *Climate Research*, 73(1–2), 135–150. <https://doi.org/10.3354/cr01465>
- Di Pierro, E. A., Mosca, E., González-Martínez, S. C., Binelli, G., Neale, D. B., & Porta, N. L. (2017). Adaptive variation in natural Alpine populations of Norway spruce (*Picea abies* [L.] Karst) at regional scale: Landscape features and altitudinal gradient effects. *Forest Ecology and Management*, 405, 350–359. <https://doi.org/10.1016/j.foreco.2017.09.045>
- Domínguez Martín, R. (2001). Las transformaciones del sector ganadero en España: (1940–1985). *Ager: Journal of Depopulation and Rural Development Studies*, 1, 47–83.
- Dullinger, S., Dirnböck, T., & Grabherr, G. (2004). Modelling climate change-driven treeline shifts: Relative effects of temperature increase, dispersal and invasibility. *Journal of Ecology*, 92(2), 241–252. <https://doi.org/10.1111/j.0022-0477.2004.00872.x>
- Escudero, A., Pérez-García, F., & Luzuriaga, A. L. (2002). Effects of light, temperature and population variability on the germination of seven Spanish pines. *Seed Science Research*, 12(4), 261–271. <https://doi.org/10.1079/SSR2002116>
- Escudero, A., Sanz, M. V., Pita, J. M., & Pérez-García, F. (1999). Probability of germination after heat treatment of native Spanish pines. *Annals of Forest Science*, 56(6), 511–520. <https://doi.org/10.1051/forest:19990608>
- Gracia, M., Meghelli, N., Comas, L., & Retana, J. (2011). Land-cover changes in and around a National Park in a mountain landscape in the Pyrenees. *Regional Environmental Change*, 11(2), 349–358. <https://doi.org/10.1007/s10113-010-0138-0>
- Grau, O., Ninot, J. M., Cornelissen, J. H. C., & Callaghan, T. V. (2013). Similar tree seedling responses to shrubs and to simulated environmental changes at Pyrenean and subarctic treelines. *Plant Ecology & Diversity*, 6(3–4), 329–342. <https://doi.org/10.1080/17550874.2013.810311>
- Greenwood, S., & Jump, A. S. (2014). Consequences of treeline shifts for the diversity and function of high altitude ecosystems. *Arctic, Antarctic, and Alpine Research*, 46(4), 829–840. <https://doi.org/10.1657/1938-4246-46.4.829>
- Harsch, M. A., Hulme, P. E., McGlone, M. S., & Duncan, R. P. (2009). Are treelines advancing? A global meta-analysis of treeline response to climate warming. *Ecology Letters*, 12(10), 1040–1049. <https://doi.org/10.1111/j.1461-0248.2009.01355.x>
- Hartig, F. (2019). *DHARMA: Residual diagnostics for hierarchical (multi-level/mixed) regression models*. R package version 0.2.4. Retrieved from <https://CRAN.R-project.org/package=DHARMA>
- Holtmeier, F.-K., & Broll, G. (2005). Sensitivity and response of northern hemisphere altitudinal and polar treelines to environmental change at landscape and local scales. *Global Ecology and Biogeography*, 14(5), 395–410. <https://doi.org/10.1111/j.1466-822X.2005.00168.x>
- Hothorn, T., Bretz, F., & Westfall, P. (2008). Simultaneous inference in general parametric models. *Biometrical Journal*, 50, 346–363. <https://doi.org/10.1002/bimj.200810425>
- IPCC (2018). Global warming of 1.5°C. In V. Masson-Delmotte, P. Zhai, H.-O. Pörtner, D. Roberts, J. Skea, P. R. Shukla et al (Eds.), *An IPCC special report on the impacts of global warming of 1.5°C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change, sustainable development, and efforts to eradicate poverty*.
- Iwaizumi, M. G., & Takahashi, M. (2012). Effects of pollen supply and quality on seed formation and maturation in *Pinus densiflora*. *Journal of Plant Research*, 125(4), 517–525. <https://doi.org/10.1007/s10265-011-0463-5>
- Iwasaki, T., Sase, T., Takeda, S., Ohsawa, T. A., Ozaki, K., Tani, N., ... Watano, Y. (2013). Extensive selfing in an endangered population of *Pinus parviflora* var. *parviflora* (Pinaceae) in the Boso Hills, Japan. *Tree Genetics & Genomes*, 9(3), 693–705. <https://doi.org/10.1007/s11295-012-0585-5>
- Johnson, J. S., Gaddis, K. D., Cairns, D. M., & Krutovsky, K. V. (2017). Seed dispersal at alpine treeline: An assessment of seed movement within the alpine treeline ecotone. *Ecosphere*, 8(1), e01649. <https://doi.org/10.1002/ecs2.1649>
- Juntunen, V., & Neuvonen, S. (2006). Natural regeneration of Scots pine and Norway spruce close to the timberline in northern Finland. *Silva Fennica*, 40, 443–458. <https://doi.org/10.14214/sf.329>
- Keshtkar, E., Mathiassen, S. K., Beffa, R., & Kudsk, P. (2017). Seed germination and seedling emergence of blackgrass (*Alopecurus myosuroides*) as affected by non-target-site herbicide resistance. *Weed Science*, 65(6), 732–742. <https://doi.org/10.1017/wsc.2017.44>
- Körner, C., & Paulsen, J. (2004). A world-wide study of high altitude treeline temperatures. *Journal of Biogeography*, 31(5), 713–732. <https://doi.org/10.1111/j.1365-2699.2003.01043.x>
- Körner, K. (2012). *Alpine treelines functional ecology of the global high elevation tree limits*. Basel, Switzerland: Springer.
- Kruse, S., Gerdes, A., Kath, N. J., Epp, L. S., Stoof-Leichsenring, K. R., Pestryakova, L. A., & Herzsich, U. (2019). Dispersal distances and migration rates at the arctic treeline in Siberia – A genetic and simulation-based study. *Biogeosciences*, 16(6), 1211–1224. <https://doi.org/10.5194/bg-16-1211-2019>
- Kullman, L. (2002). Rapid recent range-margin rise of tree and shrub species in the Swedish Scandes. *Journal of Ecology*, 90(1), 68–77. <https://doi.org/10.1046/j.0022-0477.2001.00630.x>
- Leonelli, G., Pelfini, M., di Cella, U. M., & Caravaglia, V. (2011). Climate warming and the recent treeline shift in the European Alps: The role of geomorphological factors in high-altitude sites. *Ambio*, 40, 264–273. <https://doi.org/10.1007/s13280-010-0096-2>
- Lett, S., & Dorrepaal, E. (2018). Global drivers of tree seedling establishment at alpine treelines in a changing climate. *Functional Ecology*, 32(7), 1666–1680. <https://doi.org/10.1111/1365-2435.13137>
- Lett, S., Wardle, D. A., Nilsson, M.-C., Teuber, L. M., & Dorrepaal, E. (2018). The role of bryophytes for tree seedling responses to winter climate change: Implications for the stress gradient hypothesis. *Journal of Ecology*, 106(3), 1142–1155. <https://doi.org/10.1111/1365-2745.12898>
- Liang, E., Wang, Y., Piao, S., Lu, X., Camarero, J. J., Zhu, H., ... Peñuelas, J. (2016). Species interactions slow warming-induced upward shifts of treelines on the Tibetan Plateau. *Proceedings of the National Academy of Sciences of the United States of America*, 113(16), 4380–4385. <https://doi.org/10.1073/pnas.1520582113>
- López-Moreno, J. I., & Beniston, M. (2008). Daily precipitation intensity projected for the 21st century: Seasonal changes over the Pyrenees. *Theoretical and Applied Climatology*, 95(3–4), 375–384.
- Nathan, R. (2006). Long-distance dispersal of plants. *Science*, 313(5788), 786–788. <https://doi.org/10.1126/science.1124975>
- Nathan, R., & Muller-Landau, H. C. (2000). Spatial patterns of seed dispersal, their determinants and consequences for recruitment. *Trends in Ecology & Evolution*, 15(7), 278–285. [https://doi.org/10.1016/S0169-5347\(00\)01874-7](https://doi.org/10.1016/S0169-5347(00)01874-7)
- Ninot, J. M., Batllori, E., Carrillo, E., Carreras, J., Ferré, A., & Gutiérrez, E. (2008). Timberline structure and limited tree recruitment in the Catalan Pyrenees. *Plant Ecology & Diversity*, 1(1), 47–57. <https://doi.org/10.1080/17550870802260764>
- Ninot, J. M., Carrillo, E., Font, X., Carreras, J., Ferré, A., Masalles, R. M., ... Vigo, J. (2007). Altitude zonation in the Pyrenees. A geobotanic interpretation. *Phytocoenologia*, 37(3–4), 371–398. <https://doi.org/10.1127/0340-269X/2007/0037-0371>
- Notivol, E., Oliet, J., & Serrada, R. (2012). *Pinus uncinata* Ram. In J. Pemán, R. M. Navarro, J. L. Nicolás, M. A. Prada, & R. Serrada (Eds.), *Producción y manejo de semillas y plantas forestales*. Tomo I. (pp.



- 1003–1018). Spain: Organismo Autónomo de Parques Nacionales, Ministerio de Agricultura, Alimentación y Medio Ambiente.
- Ohsawa, T., & Ide, Y. (2008). Global patterns of genetic variation in plant species along vertical and horizontal gradients on mountains. *Global Ecology and Biogeography*, 17(2), 152–163. <https://doi.org/10.1111/j.1466-8238.2007.00357.x>
- Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., & R Core Team (2019). *nlme: Linear and nonlinear mixed effects models. R package version 3.1-139*. Retrieved from <https://CRAN.R-project.org/package=nlme>
- R Core Team (2018). *R: A Language and Environment for Statistical Computing*. Vienna: R Foundation for Statistical Computing. <https://www.R-project.org>
- Rehfeldt, G. E. (1989). Ecological adaptations in Douglas-Fir (*Pseudotsuga menziesii* var. *glauca*): A synthesis. *Forest Ecology and Management*, 28(3), 203–215. [https://doi.org/10.1016/0378-1127\(89\)90004-2](https://doi.org/10.1016/0378-1127(89)90004-2)
- Ritz, C., Baty, F., Streibig, J. C., & Gerhard, D. (2015). Dose-response analysis using R. *PLoS ONE*, 10, e0146021. <https://doi.org/10.1371/journal.pone.0146021>
- Ritz, C., Pipper, C. B., & Streibig, J. C. (2013). Analysis of germination data from agricultural experiments. *European Journal of Agronomy*, 45, 1–6. <https://doi.org/10.1016/j.eja.2012.10.003>
- Šenfeldr, M., & Maděra, P. (2011). Population structure and reproductive strategy of Norway spruce (*Picea abies* L. Karst) above the former pastoral timberline in the Hrubý Jeseník Mountains, Czech Republic. *Mountain Research and Development*, 31, 131–143. <https://doi.org/10.1659/MRD-JOURNAL-D-10-00073.1>
- Sirois, L. (2000). Spatiotemporal variation in black spruce cone and seed crops along a boreal forest–Tree line transect. *Canadian Journal of Forest Research*, 30(6), 900–909. <https://doi.org/10.1139/x00-015>
- Smith, C. C., Hamrick, J. L., & Kramer, C. L. (1988). The effects of stand density on frequency of filled seeds and fecundity in lodgepole pine (*Pinus contorta* Dougl.). *Canadian Journal of Forest Research*, 18(4), 453–460. <https://doi.org/10.1139/x88-066>
- Smith, W. K., Germino, M. J., Hancock, T. E., & Johnson, D. M. (2003). Another perspective on altitudinal limits of alpine timberlines. *Tree Physiology*, 23(16), 1101–1112. <https://doi.org/10.1093/treephys/23.16.1101>
- Speed, J. D. M., Austrheim, G., Hester, A. J., & Myserud, A. (2010). Experimental evidence for herbivore limitation of the treeline. *Ecology*, 91(11), 3414–3420. <https://doi.org/10.1890/09-2300.1>
- Tapias, R., Climent, J., Pardos, J. A., & Gil, L. (2004). Life histories of Mediterranean pines. *Plant Ecology*, 171(1), 53–68. <https://doi.org/10.1023/B:VEGE.0000029383.72609.f0>
- Treml, V., Šenfeldr, M., Chuman, T., Ponocná, T., & Demková, K. (2016). Twentieth century treeline ecotone advance in the Sudetes Mountains (Central Europe) was induced by agricultural land abandonment rather than climate change. *Journal of Vegetation Science*, 27(6), 1209–1221. <https://doi.org/10.1111/jvs.12448>
- Viechtbauer, W. (2010). Conducting meta-analyses in R with the metafor package. *Journal of Statistical Software*, 36, 1–48.
- Vitali, A., Camarero, J. J., Garbarino, M., Piermattei, A., & Urbinati, C. (2017). Deconstructing human-shaped treelines: Microsite topography and distance to seed source control *Pinus nigra* colonization of treeless areas in the Italian Apennines. *Forest Ecology and Management*, 406, 37–45. <https://doi.org/10.1016/j.foreco.2017.10.004>
- Vitali, A., Garbarino, M., Camarero, J. J., Malandra, F., Toromani, E., Spalevic, V., ... Urbinati, C. (2019). Pine recolonization dynamics in Mediterranean human-disturbed treeline ecotones. *Forest Ecology and Management*, 435, 28–37. <https://doi.org/10.1016/j.foreco.2018.12.039>
- Wickham, H. (2016). *ggplot2: Elegant graphics for data analysis*. New York: Springer.
- Zuur, A. F., Ieno, E. N., Walker, N. J., Saveliev, A. A., & Smith, G. M. (2009). *Mixed effects models and extensions in ecology with R*. New York: Springer.

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section.

Appendix S1. Altitude and coordinates of the sampling plots for the seed rain experiment.

Appendix S2. Climate data for meteorological stations spanning the geographical area where our study sites are located.

Appendix S3. Assessment of cone production at the two extremes of the elevation gradient.

Appendix S4. Analysis of the relationship between seed rain and the density and diameter at breast height (DBH) of reproductive trees.

Appendix S5. Correlations between the number of full seeds and cone length at each elevation position.

Appendix S6. Proportion of seeds germinated out of the seeds sown at each elevation position and study site.

Appendix S7. Germination dynamics (percentage of cumulative germination) for each site and elevation provenance.

Appendix S8. Results of the meta-analytic random-effects models on germination dynamics for the three estimated parameters (t_{50} ; b ; d).

Appendix S9. Seeds recovered in the seed traps at each elevation position and study year.

How to cite this article: Anadon-Rosell A, Talavera M, Ninot JM, Carrillo E, Batllori E. Seed production and dispersal limit treeline advance in the Pyrenees. *J Veg Sci*. 2020;31:981–994. <https://doi.org/10.1111/jvs.12849>